

COST Action: ES1306

COST STSM Reference Number: COST-STSM-ES1306-27734

Effect of alternative soil management in Mediterranean vineyards to reduce sediment and Carbon losses due to water erosion

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FINAL REPORT

Because it reduces soil fertility, damages nearby roads, and causes floods, soil erosion is an important problem in vineyards (Lieskovský and Kenderessy, 2014; Martínez Casanovas, 2015). Higher erosion rates have been recorded in vineyards than in other land use areas (Cerdan et al., 2010) in the Mediterranean region (Vanmaerke et al., 2011) because of several characteristics of the soil, soil management, and climate (Novara et al., 2011; Tarolli et al., 2015). First, the soil in traditional Mediterranean vineyards is bare for most of the year; the cover is only significant during the summer, when there is almost no rain other than sparse and irregular storms. Bare soils result in high erosion rates, and a recovery of the vegetation contribute to an important reduction of the soil and nutrient losses in the Mediterranean region (Cerdà, 1998; Novara et al., 2013, 2015) and Africa (Mekkonen et al., 2015). Second, traditional soil management in Mediterranean vineyards includes continuous tillage with the goal of eliminating competition between vines and other plants for water and nutrients. Although tillage also reduces evaporation in the Mediterranean region, it results in high erosion rates (García Orenes et al., 2012). Third, a high soil organic matter content can help reduce erosion, but soils of the Mediterranean vineyards have low organic matter content because of the low inputs of organic matter and because the climate promotes high mineralization rates. The organic matter content is further reduced by the tillage. Fourth, vineyard soils in many regions are shallow and stony, which increases their vulnerability to soil erosion. Finally, the large vine-producing regions in Mediterranean region are hilly and experience high intensity rainfall events, both of which will obviously increase the potential for erosion.

Within this context, alternative management, such as reduced tillage, the application of mulch or organic fertilizer, or the planting of cover crops, have been developed to protect soil from erosion. These alternative management methods generally increase the input of soil organic matter (SOM). The importance of SOM in reducing soil erosion is well known, i.e., SOM reduces erosion by improving soil structure and aggregate stability (Balesdent et al., 2000; Barthès and Roose, 2002; Six et al., 2004). Erosion rates are lower and SOM contents are higher in vineyards that are planted with cover crops and are not tilled than in vineyard that are managed with bare soil and traditional tillage (Biddoccu et al., 2014; Ruiz-Colmenero et al., 2013; Virto et al., 2012). Hence, the adoption of soil-conservation practices is encouraged both to prevent erosion and to sequester atmospheric carbon dioxide (CO₂) in the soils.

Although agricultural conservation practices, as well as the planting of cover crops are included in the Agro-environmental management (AEM), few studies have analyzed how the carbon (C) cycle and erosion processes are altered in vineyards with alternative soil management. The apparently positive effect of higher soil C content on soil properties and climate change mitigation could lead to C loss in the system in terms of higher CO₂ emissions or higher amounts of C in sediments, i.e., in soil that is transported from the field by erosion (Bienes et al., 2010). Gao et al. (2012, 2013) introduced the C health threshold theory, which indicates that increases in soil C levels could lead to ecosystem degradation. Gao et al. also determined that if C storage exceeds nutrient and water supply limits, an ecosystem will fall into a sub-health state of fitness; after that, C will be lost through soil erosion or other pathways. In the current study, the C threshold is defined as that level of SOC in an AEM managed vineyard above which erosion will result in a greater loss of C than occurs in a comparable vineyard that is managed with conventional tillage (CT). Several authors have studied the C health threshold theory with respect to afforested and natural soils (Gao et al., 2012; Wang and Cao, 2011), but the theory has not been investigated in a semiarid cultivated soil under erosion processes.

The aim of the STS mission was to investigate the effect of different soil management in vineyards on soil organic matter content and erosion losses, comparing two Mediterranean regions (Italy and Spain). The study attempted to answer three questions with respect to vineyards located on hillsides in the semi-arid environment of the Mediterranean region: i) Is increasing the SOC always desirable?; ii) Can the SOC stock be increased under AEM without resulting in high C loss due to erosion? and iii) Is the C threshold measurable?

1.1 The C threshold concept

Erosion results in C loss via three major pathways: (i) C contained in soil that is transported and deposited elsewhere as sediment; (ii) dissolved organic carbon (DOC) contained in runoff; and (iii) CO₂ emission (Jachinte and Lal, 2001). Among these pathways, the first is most important, i.e., most C is lost as SOC in sediment. The other two pathways, although relevant for the global C budget and for ecological properties, are dependent on sediment transport and C content. Using data for SOC stocks and dynamics from long-term experiments in different regions, Jachinte and Lal (2001) found that erosion-induced CO₂ emission rates ranged from 6 to 52 g C m⁻² yr⁻¹. Similarly, Van Hemelryck et al. (2010) estimated that soil redistribution processes resulted in an additional loss of 2 to 12% of C from eroded sediment via CO₂ emission. Both studies showed that erosion induced-CO₂ emission depends on the C content of the soil and sediment. Similarly, the DOC in runoff represents a low percentage of the total C loss (Mchunu and Chaplot, 2012). It follows that

the quantity of C lost during erosion can be reasonably estimated from the quantity of SOC that is transported with eroded soil and that is deposited elsewhere as sediment.

The loss of C in soil sediments ($C_{loss_{sediment}}$) can be described by the following linear relationship (Starr et al., 2000):

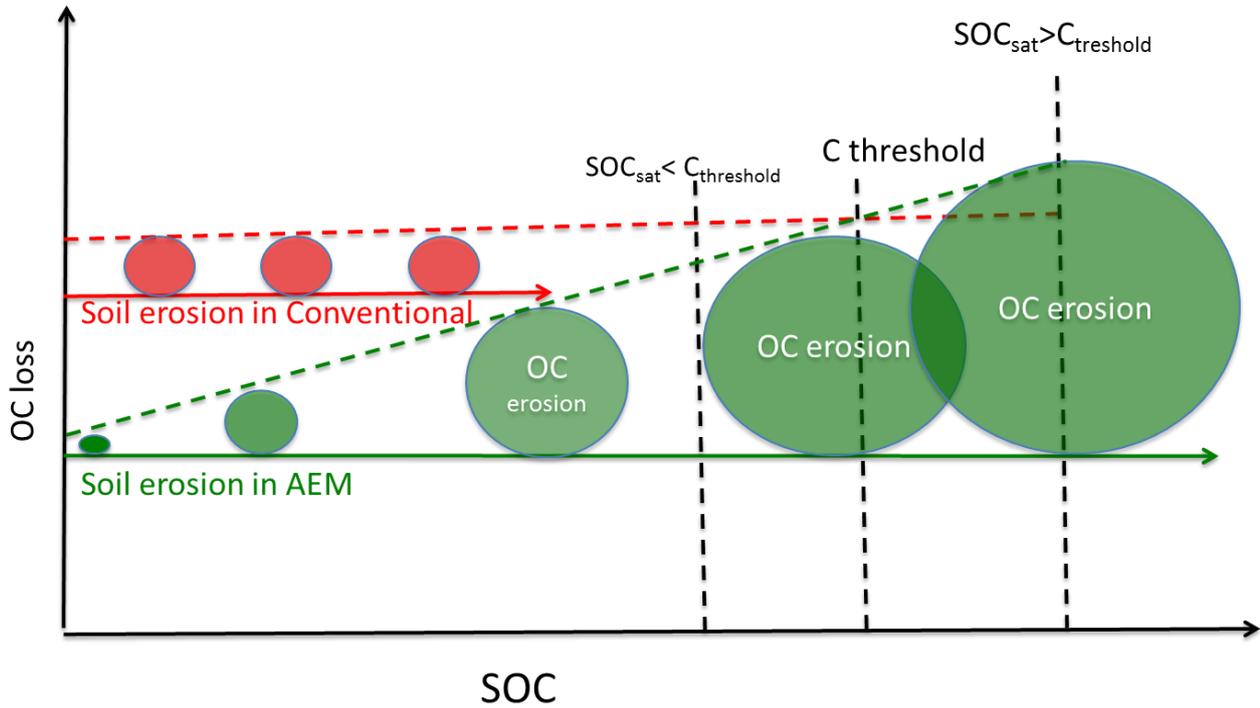
$$C_{loss_{sediment}} = SE * SOC * E_r \quad (\text{Eq. 1})$$

where SOC is the content of organic C in soil (%), E_r is the enrichment ratio of eroded sediment relative to the original soil (dimensionless), and SE is soil erosion rate ($\text{Mg ha}^{-1}\text{y}^{-1}$). According to equation 1, C loss increases with the erosion rate and SOC content. SOC and SE are both functions of organic matter input into the soil, i.e., increases in soil organic matter increase SOC and reduce SE because organic matter increases soil aggregate stability (Loveland and Webb, 2003).

In sloping vineyards, alternative soil management (AEM management, i.e., management without tillage and with a covercrop) reduces erosion relative to conventional tillage (CT) because the non-vine vegetation reduces the impact of rain drops on the soil, produces biomass that contributes to increases in SOC (Novara et al., 2011) and therefore to aggregate stability (Blavet et al., 2009). The higher SOC level resulting from continuous AEM management, however, produces C-enriched sediments and consequently could lead to higher C losses than with CT, despite the lower SE (Figure 1). Considering that possibility and as noted earlier, we define the SOC threshold as that level of SOC under AEM management that results in a C loss with AEM management that is equal to the C loss under CT management ($C_{loss_{CT}}$) (Figure 1).

If the soil C saturation level (the maximum, steady state level of C that can accumulate in a specific soil) is higher than the SOC threshold, the C threshold will correspond to a C_{AEM} value; if the soil C saturation level is lower than the C threshold, the C threshold will be equivalent to the C saturation value. We indicated C saturation level (or C steady state) the maximum level of C accumulated in a certain soil, despite the C input increasing.

Figure 1. Conceptual diagram of dynamical change in C loss in sloping area with different soil management. The soil erosion is lower under AEM management (green line) than CT management (red line). The increase of SOC in AEM management (green circle) can entail an higher OC loss in AEM than CT. The cross point between dotted red line (OC loss under CT) and dotted green line (OC loss under AEM) is described as C threshold in sloping area.



Considering constant for both soil managements the environmental conditions, the SOC threshold is calculated with the following equations (2 and 3):

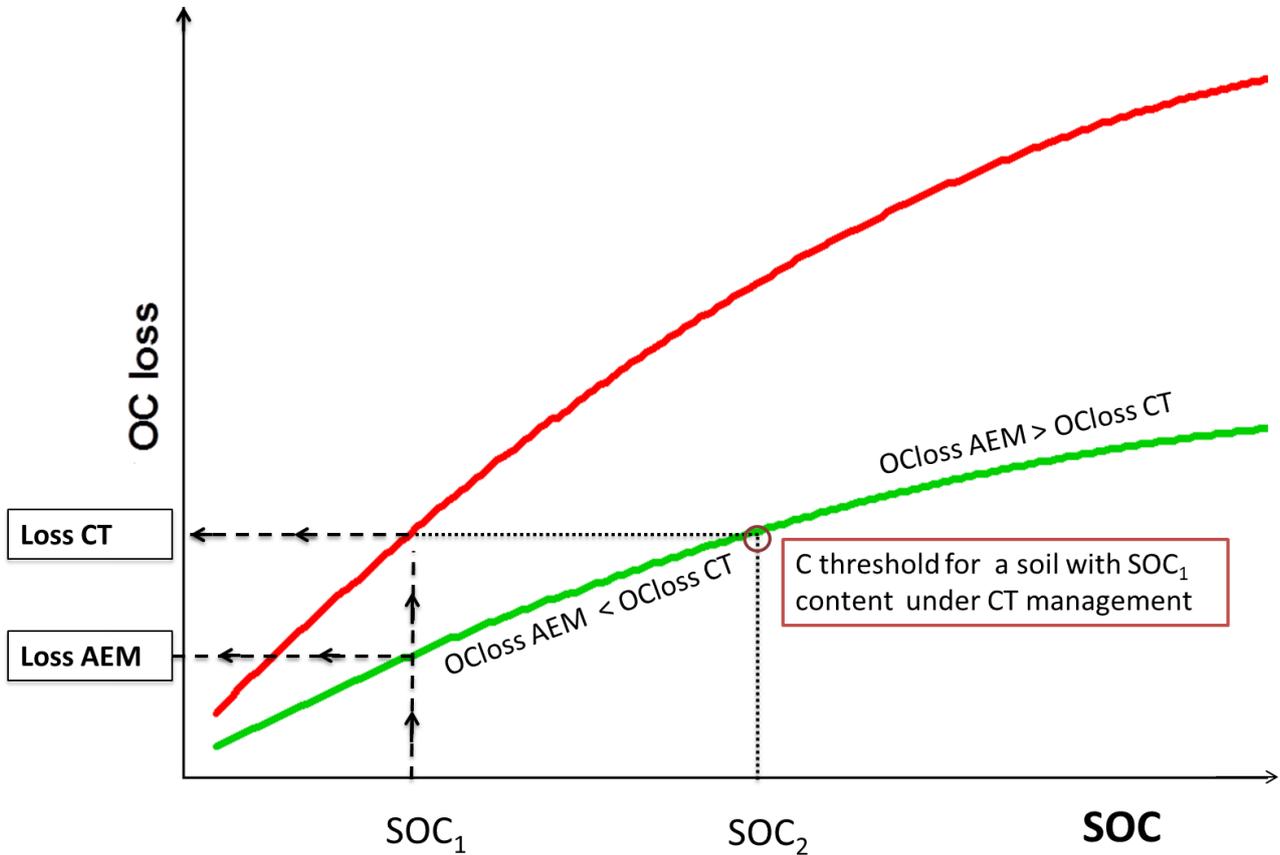
$$Closs_{AEM} = Closs_{CT} \quad (\text{Eq. 2})$$

and according to equation 1, it follows that:

$$SE_{AEM} * SOC_{AEM} * E_{rAEM} = SE_{CT} * SOC_{CT} * E_{rCT} \quad (\text{Eq. 3})$$

Based on equation 3, the C losses relative to SOC content with AEM and CT management are presented in Figure 2 (green and red lines). Considering the same value of SOC (C_1) for CT and AEM, $Closs$ will be higher for CT ($Closs_{CT}$) than AEM ($Closs_{AEM}$), given that the erosion rate SE will be higher in CT than in AEM because of differences in soil cover.

Figure 2. C loss under two different soil managements. The black circle indicates the SOC threshold.



If no disturbance occurs, the SOC content in CT can be considered constant over many years. Given a value of SOC under CT (SOC_1) with $C_{loss_{CT}}$, the $C_{loss_{AEM}}$ will be reached with a SOC content equal to SOC_2 . Values higher to SOC_2 will determine a higher C_{loss} in AEM than in CT. The SOC_2 value can, therefore, be considered the C threshold for a given soil that, if managed under CT, will contain a steady state level of SOC equal to SOC_1 (Figure 2).

2. Materials and Methods

2.1 Study area and soil sampling

The study areas were located in southern Sicily (IT) and in Valencia province (SP). Vineyards in Sicily and in Valencia are commonly managed with CT (at least five shallow tillages per year) to control weeds and reduce water competition. Recently, alternative soil management in vineyards is spreading thanks to AEM. In particular, AEM management in vineyards involves

annual cover cropping using legumes like faba bean (*Vicia faba*) and vetch (*Vicia sativa*); after the covercrops are disked into the soil in spring, the vineyard is subjected to two shallow tillages.

In the study areas, 95 paired sites were chosen (10 paired sites in Valencia and 85 paired site in Sicily). The paired-site approach was used to compare SOC stocks after 5 or more years of management with AEM vs. CT (Novara et al., 2012). The plots at each pair of sites (one plot per site) were similar with respect to soil type, slope, elevation, exposure, and drainage.

The AEM applied in Sicilian paired sites was Cover crop soil management with Vicia Faba; the AEM applied in Valencia were Cover cropping with Vicia sativa and application of organic manure. Three soil samples were collected at 0-15 cm depth in each plot. The soil was dried and passed through a 2-mm sieve before SOC was quantified according to Walkley and Black (1934), The analysis were carried out in the Department of Geography, Valencia University.

2.2 SOC threshold calculation

The SOC threshold for each pair of sites was calculated according to equation 3. SE_{AEM} and SE_{CT} were estimated using the USLE equation (Wischmeier, 1978):

$$SE (Mgha^{-1}) = K * R * C * LS * P \quad (\text{eq. 4})$$

where K is the soil erodibility, C is C_{factor} , R is the rainfall erosivity, LS is a topographic factor, and P is support practice.

Because R, LS, and P factors were the same for the two plots of each pair of sites, the calculation of SE was simplified by their exclusion.

Soil erodibility (K) for each plot of paired site was calculated based on texture, organic matter content, and permeability according to equations presented by Wishmeier and Smith (1971) and Renard et al. (1997); these equations are recommended when the organic matter content is known and when the silt content is < 70%. Values for the C_{factor} were 0.65 in CT plots and 0.22 in AEM plots (Novara et al., 2011). The same value was used for E_{rAEM} and E_{rCT} . Ruiz-Colmenero et al. (2013) found different values of SOC in sediment and soil but similar ratios for vineyards managed with conventional tillage ($E_r = 1.4$) and with a *Secale cereale* cover crop ($E_r = 1.5$).

The equation 3 was arranged as follows:

$$K_{AEM} * SOC_{AEM} = \frac{K_{CT} * SOC_{CT} * E_{rCT}}{E_{rAEM}} \quad (\text{Eq. 5})$$

as:

$$K=f(\text{SOC}_{\text{AEM}}), \quad (\text{Eq. 6})$$

follows:

$$\text{SOC}_{\text{threshold}} = \text{SOC}_{\text{AEM}} = \frac{K_{\text{CT}} * \text{SOC}_{\text{CT}} * E_{\text{rCT}}}{E_{\text{rAEM}} * f^{-1}} \quad (\text{Eq. 7})$$

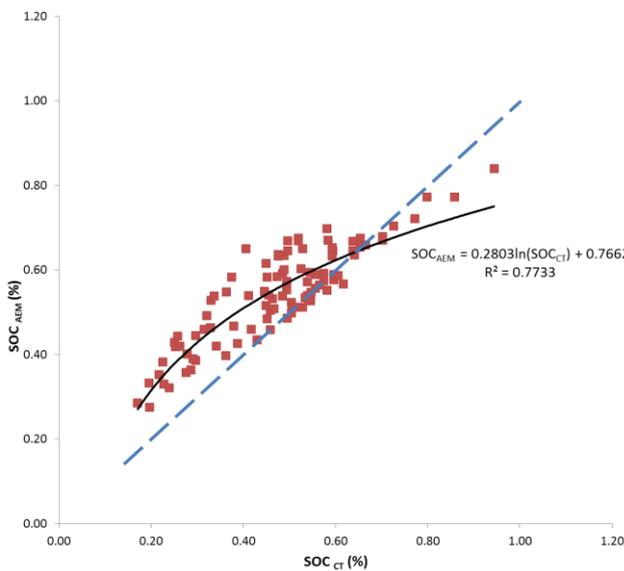
The mean and standard deviation of the SOC values were calculated for each plot.

3. Results and discussion

3.1 SOC as affected by CT and AEM soil management

After AEM management was initiated the C increase was moderate (Figure 3). Among the 95 AEM plots, the highest SOC gain was 0.24%, and the average was 0.046% . The higher C sequestration rates in AEM plots occurred in those pairs with low SOC values in CT plots; the sequestration rate dropped as the SOC in CT plots approached 0.66%. With this high SOC value in CT plots, the C sequestration rate was 0, and 0.66% SOC content was therefore assumed to be the C saturation level for vineyard soils given the environmental conditions of the study area and the soil management performed.

Figure 3. Soil organic carbon (SOC%) in each paired site for the two soil managements. The dotted blu line represents SOC under Conventional tillage (CT). The crossing point between the logarithmic curve (black line) and linear equation (dotted line) describes the C saturation level.

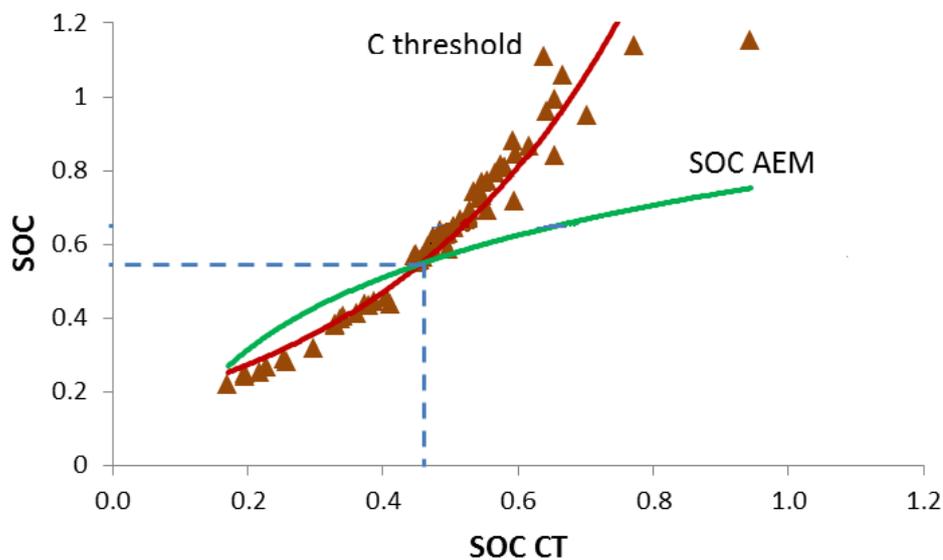


3.2 SOC threshold

The SOC threshold, which was calculated for each paired site according to equation 7, showed an exponential trend (Figure 4). In the absence of limiting factors, which reduce the capacity of soil to sequester C, the increase in organic C exponentially improves soil structure and soil chemical properties, leading to a lower erosion rate and soil C sink ability. Soil erodibility and consequently C_{loss} increased with the silt content of soil (Perez-Rodriguez et al., 2007). Soils with high silt are more susceptible to erosion and have a reduced ability to retain C, leading to lower SOC threshold values. After AEM adoption, SOC_{AEM} exceeded the SOC threshold in soils in which the CT plots had a low SOC content. For values of SOC under CT ranging from 0.2% to 0.45%, the SOC_{AEM} reached the SOC threshold, and in these cases the values for C_{loss} were higher with AEM than with CT.

Similar results were found by Ruiz-Colmenero et al. (2013), who compared SOC in a vineyard managed with a cover crop vs. CT. The latter researchers reported that less sediment was generated with a cover crop than with CT but that the sediment from the cover crop soil contained about 1.4-times more SOC than the sediment from the CT soil. For values of SOC ranging from 0.45% to 0.66% (the saturation level) with CT management in the current study, the SOC threshold was higher than SOC_{AEM} . In these cases, the adoption of AEM should be encouraged up to reach the SOC threshold.

Figure 4. SOC threshold and measured SOC after AEM adoption



4. Implications and Conclusion

AEM management using cover crops in vineyards increases SOC because it increases C inputs and reduces soil organic matter mineralization, such that there is a net increase in C stock with AEM management relative to CT management. In sloping areas, however, selecting the best AEM depends on an understanding of C loss resulting from water erosion. Our results showed that in some cases (high C input through organic fertilization) the loss of C is greater with AEM management than with CT management. In these cases, the difference between SOC_{AEM} and the SOC threshold indicates the increase in C loss resulting from AEM adoption; the difference between the SOC threshold and SOC_{CT} , on the other hand, indicates the potential of the soil to increase its SOC stock while maintaining the same C losses with AEM management. In other cases, the SOC threshold was higher than AEM; in these cases, the difference between the SOC threshold and SOC_{AEM} represents the potential quantity of C that could be sequestered by soil.

To maximize the effectiveness of AEM management both for the environment and for the optimization of SOC input, European policies should consider many factors such as the pedoclimate, cover crop management, plant species, and the quality and quantity of cover crop biomass. Knowledge about a soil's potential to sequester C or different scenario of AEM adoption is important for designing policy concerning vineyard soil management. This paper presents a method to evaluate the effects of AEM management to reduce erosion losses on both the increase in SOC and the risk of C loss. As consequence, this paper could provide a tool for payment diversification in relation to agro-ecosystem services provided by AEM management.

Reference

- Balesdent, J., Chenu, C., Balabane, M., 2000. Relationship of organic matter dynamics to physical protection and tillage. *Soil and Tillage Research*. 53, 215-230.
- Barthès, B., Roose, E., 2002. Aggregate stability as an indicator of soil susceptibility to runoff and erosion; validation at several levels. *Catena* 47, 133-149.
- Batjes, N.H., 2014. Projected changes in soil organic carbon stocks upon adoption of recommended soil and water practices in the Upper Tana River Catchment, Kenia. *Land Degradation and Development*, 25, 278–287. DOI: 10.1002/ldr.2141

- Biddoccu, M., Opsi, F., and Cavallo, E., 2014. Relationship between runoff and soil losses with rainfall characteristics and long-term soil management practices in a hilly vineyard (Piedmont, NW Italy). *Soil Science and Plant Nutrition*, 60, 92-99.
- Blavet, D., DeNoni, G., LeBissonnais, Y., Leonard, M., Maillo, L., Laurent, J.Y., Asseline, J., Leprun, J.C., Arshad, M.A., Roose, E., 2009. Effect of land use and management on the early stages of soil water erosion in French Mediterranean vineyards. *Soil and Tillage Research* 106, 124–136. 26652528
- Cerdà, A., 1998. The influence of aspect and vegetation on seasonal changes in erosion under rainfall simulation on a clay soil in Spain. *Canadian Journal of Soil Science* 78, 321-330.
- Cerdan, O., Govers, G., Le Bissonnais, Y., Van Oost, K., Poesen, J., Saby, N., Gobin, A., Vacca, A., Quinton, J., Auerswald, K., Klik, A., Kwaad, F.J.P.M., Raclot, D., Ionita, I., Rejman, J., Rousseva, S., Muxart, T., Roxo, M.J., Dostal, T., 2010. Rates and spatial variations of soil erosion in Europe: A study based on erosion plot data, *Geomorphology* 122 (1–2), 167-177.
- Debasish, S., Kukal, S.S., Bawa, S.S., 2014. Soil organic carbon stock and fractions in relation to land use and soil depth in the degraded Shiwalik hills of lower Himalayas. *Land Degradation and Development* 25, 407–416. DOI: 10.1002/ldr.2151.
- Gao, Y., Yu, G., He, N., 2013. Equilibration of the terrestrial water, nitrogen, and carbon cycles: Advocating a health threshold for carbon storage. *Ecological Engineering* 57, 366-374.
- Gao, Y., Yu, G.R., He, N.P., He, H.L., Wang, Q.F., Fang, H.J., 2012. Is there an existing healthy threshold for carbon storage in the ecosystem? *Environ. Sci. Technol.* 46 (9), 4687–4688.
- García-Orenes, F., Roldán, A., Mataix-Solera, J., Cerdà, A., Campoy, M., Arcenegui, V., Caravaca, F., 2012. Soil structural stability and erosion rates influenced by agricultural management practices in a semi-arid Mediterranean agro-ecosystem. *Soil Use and Management* 28(4), 571-579. DOI: 10.1111/j.1475-2743.2012.00451.x
- Li, Q.J., Fang, H.Y., Sun, L. Y., Cai, Q.G., 2014. Using the ¹³⁷Cs technique to study the effect of soil redistribution on soil organic carbon and total nitrogen stocks in an agricultural catchment of

- northeast China. *Land Degradation & Development* 25, 350–359 DOI: 10.1002/ldr.2144
- Lieskovský J, Kenderessy. P., 2014. Modelling the effect of vegetation cover and different tillage practices on soil erosion in vineyards: a case study en Vráble (Slovakia) using WATEM/SEDEM. *Land Degradation and Development* 25, 288-296. DOI: 10.1002/ldr.2162.2014.
- Loveland, P., Webb, J., 2003. Is there a critical level of organic matter in the agricultural soils of temperate regions: a review *Soil & Tillage Research*. 70, 1-18
- Lozano-García, B., Parras-Alcántara, L., 2014. Variations in soil organic carbón and nitrogen stocks along a toposequence in a traditional mediterranean olive grove. *Land Degradation and development*, 25, 297–304 | DOI: 10.1002/ldr.2284
- McHunu, C., Chaplot, V., 2012. Land degradation impact on soil carbon losses through water erosion and CO2 emissions. *Geoderma* 177, 72-79.
- Martínez-Casasnovas, J. A., Ramos, M. C., Benites, G., 2015. Soil and water assessment tool soil loss simulation in the sub-basin scale in the Alt Penedès-Anoia vineyard region (NE Spain) in the 2000s. *Land Degradation & Development*.| DOI: 10.1002/ldr.2240
- Mekonnen, M., Keesstra, S. D., Stroosnijder, L., Baartman, J. E., Maroulis, J., 2015. Soil conservation through sediment trapping: A review. *Land Degradation & Development*. DOI: 10.1002/ldr.2308
- Jaiarree, S., Chidthaisong, A., Tangtham, N., Polprasert, C., Sarobol, E., Tyler, S.C., 2014. CarbonBudget and sequestration potential in a sandy soil treated with compost. *Land Degradation & Development* 25, 120-129.
- Novara, A., La Mantial, T., Barbera, V., Gristina, L., 2012. Paired-site approach for studying soil organic carbon dynamics in a Mediterranean semiarid environment *Catena* 89 (1), 1-7.
- Novara, A., Gristina, L., Guitoli, F., Santoro, A., Cerdà, A., 2013. Managing soil nitrate with cover crops and buffer strips in Sicilian vineyards. *Solid Earth* 4, 255-262, doi:10.5194/se-4-255-2013.

- Novara, A., Gristina, L., Saladino, S. S., Santoro, A., and Cerdà, A., 2011. Soil erosion assessment on tillage and alternative soil managements in a Sicilian vineyard. *Soil and Tillage Research* 117, 140-147.
- Novara A., Cerdà A., Dazzi C., Lo Papa G., Santoro A., Gristina L., 2015. Effectiveness of carbon isotopic signature for estimating soil erosion and deposition rates in Sicilian vineyards. *Soil tillage and Research* 152, 1-7.
- Pérez-Rodríguez, R., Marqués, M.J., Bienes, R., 2007. Spatial variability of the soil erodibility parameters and their relation with the soil map at subgroup level. *Science of the Total Environment* 378, 166-173.
- Renard, K.G., Foster, G.R., Weessies, G.A., McCool, D.K., 1997. Predicting soil erosion by water: a guide to conservation planning with the Revised Universal Soil Loss Equation (RUSLE). In: Yoder DC, editor. U.S. Department of Agriculture, Agriculture Handbook 703; 1997
- Ruiz-Colmenero, M., Bienes, D.J., Marques, M.J., 2011. Soil and water conservation dilemmas associated with the use of green cover in steep vineyards. *Soil and Tillage Research* 117, 211-223.
- Ruiz-Colmenero, M., Bienes, R., Eldridge, D.J., Marques, M.J., 2013. Vegetation cover reduces erosion and enhances soil organic carbon in a vineyard in the central Spain. *CATENA* 104, 153-160.
- Six, J., Bossuyt, H., Degryze, S., Denef, K., 2004. A history of research on the link between (micro) aggregates, soil biota, and soil organic matter dynamics. *Soil and Tillage Research* 79, 7-31.
- Starr, G.C., Lal, R., Malone, R., Hothem, D., Owens, L., Kimble, J., 2000. Modeling soil carbon transported by water erosion processes. *Land Degrad. Develop.* 11, 83–91.
- Tarolli, P., Sofia, G., Calligaro, S., Prosdocimi, M., Preti, F., Dalla Fontana, G., 2015. Vineyards in terraced landscapes: new opportunities from lidar data. *Land Degradation & Development* 26(1), 92-102.
- Vanmaercke, M., Poesen, J., Verstraeten, G., De Vewnte, J., Ocakoglu, F., 2011. Sediment yield in Europe: spatial patterns and scale dependency. *Geomorphology* 130, 142–161.

- Van Hemelryck, H., Fiener, P., Van Oost, K., Govers, G., Merckx, R., 2010. The effect of soil redistribution on soil organic carbon: an experimental study. *Biogeosciences* 7, 3971–3986. doi: 10.5194/bg-7-3971-2010
- Virto, I., Imaz, M.J., Fernández-Ugalde, O., Urrutia, I., Enrique, A., Bescansa, P., 2012. Soil quality evaluation following the implementation of permanent cover crops in semi-arid vineyards. Organic matter, physical and biological soil properties. *Spanish Journal of Agricultural Research*. 10(4), 1121-1132.
- Walkley, A., Black, A.I., 1934. An examination of Degtjareff method for determining soil organic matter and a proposed modification of the chromic acid titration method. *Soil Science* 37, 29–38.
- Wang, Y., Cao, S., 2011. Carbon sequestration may have negative impacts on ecosystem health, *Environ. Sci. Technol.*, 45, 1759–1760.
- West, T.O., Six, J., 2007. Considering the influence of sequestration duration and carbon saturation on estimates of soil carbon capacity. *Climate Change* 80, 25-41.
- Wischmeier, W., Johnson, C., Cross, B., 1971. A soil erodibility nomograph for farmland and construction sites. *J Soil Water Conserv*, 26(3),189–93.
- Wischmeier, W., Smith, D., 1978. Predicting rainfall erosion losses: a guide to conservation planning. *Agricultural Handbook No. 537*. Washington DC, USA: U.S. Department of Agriculture; 1978.